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Etymology of Ghoti

George Bernard Shaw (1856–1950), polymath, playwright, Nobel prize winner, and the most prolific letter writer in history, was an advocate of English spelling reform. He was reportedly fond of pointing out its absurdities by proving that 'fish' could be spelt 'ghoti'. That is: 'gh' as in 'rough', 'o' as in 'women' and 'ti' as in palatial.

The seascape nursery: a novel spatial approach to identify and manage nurseries for coastal marine fauna

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Abstract

Coastal marine and estuarine ecosystems are highly productive and serve a nursery function for important fisheries species. They also suffer some of the highest rates of degradation from human impacts of any ecosystems. Identifying and valuing nursery habitats is a critical part of their conservation, but current assessment practices typically take a static approach by considering habitats as individual and homogeneous entities. Here, we review current definitions of nursery habitat and propose a novel approach for assigning nursery areas for mobile fauna that incorporates critical ecological habitat linkages. We introduce the term 'seascape nurseries', which conceptualizes a nursery as a spatially explicit seascape consisting of multiple mosaics of habitat patches that are functionally connected. Hotspots of animal abundances/productivity identify the core area of a habitat mosaic, which is spatially constrained by the home ranges of its occupants. Migration pathways connecting such hotspots at larger spatial and temporal scales, through ontogenetic habitat shifts or inshore–offshore migrations, should be identified and incorporated. The proposed approach provides a realistic step forward in the identification and management of critical coastal areas, especially in situations where large habi-

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tat units or entire water bodies cannot be protected as a whole due to socio-economic, practical or other considerations.

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Introduction

Coastal ecosystems provide a range of valuable ecosystem services, such as fisheries production, protection against coastline erosion and carbon sequestration (Costanza *et al.* 1997). With about 60% of the world's population living within 100 km of the coastline (Vitousek *et al.* 1997), these ecosystems have suffered from rapid degradation (Waycott *et al.* 2009). Coastal and estuarine systems are highly productive and important for food security and livelihoods. Where multiple ecosystems are hydrologically and ecologically connected, a key function is the replenishment of offshore populations of commercially and ecologically important species of fish and crustaceans (Beck *et al.* 2001). The nursery function of these systems has received much attention over the last decade, but current procedures for identifying and evaluating critical habitats lag our scientific understanding of processes that drive nursery function and productivity. In this perspective, we propose a novel approach for delineating nursery areas for mobile fauna, incorporating ecological habitat linkages resulting from animal movements that occur at different spatial and temporal scales.

Three lines of research tackle the issue of coastal ecosystem connectivity for marine fauna, but at different conceptual scales. Firstly, the nursery-role hypothesis is mainly focused on identifying the nursery habitats that contribute most to offshore adult populations (Beck *et al.* 2001; Nagelkerken 2009). Secondly, ecosystem-connectivity studies have largely attempted to correlate a variety of structural metrics of coastal nursery habitats to catches of offshore fishery stocks (Manson *et al.* 2005). Finally, seascape studies have applied techniques and concepts from landscape ecology to understand what drives the spatial patterning of animal communities in coastal nursery habitats (Sheaves and Johnston 2008; Boström *et al.* 2011).

While each of these research directions has received increasing attention in the last decade or two, lack of integration between them has led to gaps in the development of appropriate conservation and management strategies.

The nursery-role and ecosystem-connectivity approaches typically consider critical habitats as individual, homogeneous entities. This potentially forces managers faced with conflicting objectives for conservation and alternative uses to evaluate and then trade off entire habitats against one another when determining priorities (Weinstein 2008). Moreover, protected areas with fixed boundaries are ineffective in protecting moving or transient species (Rayfield *et al.* 2008). The seascape-ecology approach points to a different solution, based on mosaics of habitat patches at smaller spatial scales (Simenstad *et al.* 2000). The spatial characteristics of habitat patches play an important role in structuring associated animal communities, but typically are not considered in assessments of nursery value, leaving a critical knowledge and conservation gap (Beck *et al.* 2001; Adams *et al.* 2006; Boström *et al.* 2011).

Previous attempts to define marine nurseries have provided an important, but relatively basic, framework for the identification of nursery habitats. These approaches are static in that they do not indicate how to specifically incorporate dynamic processes, such as ontogenetic habitat shifts, animal movement and spatially explicit usage of habitat patches and corridors within seascapes. This static, single-habitat approach potentially leads to incomplete or incorrect identification of critical habitats. The aim of this study is to take a more holistic approach in identifying nurseries. We view a nursery as a spatially explicit seascape unit (rather than a habitat unit) consisting of functionally connected mosaics of habitats incorporating ecological processes driven by animal behaviour, and define this as the 'seascape nursery'.

Review of nursery-function definitions

Early designations of nursery habitats simply referred to habitats with high densities of juvenile animals. Beck *et al.* (2001) greatly improved the definition by arguing that a nursery is a habitat contributing a higher than average biomass of juveniles per unit area to the adult population than other habitats, resulting from higher densities, higher growth, lower mortality and/or greater movement. However, this approach under-appreciates juvenile habitats that have a large surface area but low density of organisms, even though their overall contribution to the adult population might be larger. Therefore, Dahlgren *et al.* (2006) suggested that identification of nurseries should be based on their total contribution to the adult population. This was criticized as an approach that failed to consider the importance of dynamic processes that underpin nursery function (Sheaves *et al.* 2006), but no specific solutions were offered (Layman *et al.* 2006). While some studies (e.g. Beck *et al.* 2001; Adams *et al.* 2006) have covered important factors that regulate nursery value, no significant steps towards a more comprehensive and realistic method for the identification of nurseries have occurred. Clearly, managing a nursery habitat as a whole unit will not be effective without considering the sequence of habitats that are used throughout ontogeny, while other aspects of nursery habitats (e.g. movement corridors, density hotspots) should be considered to conserve the most productive and important habitat patches within nursery habitats. Some of these aspects have been briefly mentioned in previous studies (Beck *et al.* 2001; Adams *et al.* 2006), but a framework of how to address these issues is still lacking. In the present study, we propose a potential framework to enhance identification and conservation of nurseries.

We concur with the current view that the value of nurseries (as defined by Beck *et al.* 2001) relates to their ultimate contribution to the support of populations. However, we move beyond the approaches that identify nurseries as static habitat units and provide a perspective on how advances in seascape ecology can enhance designation and valuation of nursery habitats for animals that use inshore habitats before migrating offshore ('ontogenetic shifters'; Adams *et al.* 2006). Like previous efforts, our goal is to

improve the management and conservation of critical nursery habitats. Here, we build on those efforts to gain an improved measure for nursery habitat designation that captures critical processes and habitat linkages that underpin nursery function and might otherwise be missed by earlier approaches.

Early-juvenile population bottlenecks: identifying critical settlement habitats

Searching for preferred habitat while in the water column is risky and therefore settlement-stage larvae often occupy the first-encountered suitable habitat when entering estuaries or lagoons from the open ocean (Grol *et al.* 2011), with subsequent shifts to other habitats in a stepwise pattern (Cocheret de la Morinière *et al.* 2002). Less structurally complex habitats such as sand patches, macroalgal clumps or dead coral rubble may function as important settlement habitats (Dahlgren and Eggleston 2000), but are often disregarded in their value for settling larvae. The identity of transient settlement habitats is unknown for many species, they may be occupied only briefly, yet they may well form population bottlenecks for early post-settlement stages (Fodrie *et al.* 2009). They are easily missed because of the small sizes at which juveniles occupy these transient habitats and because of the relatively short duration of occupancy. However, many species settle from the plankton during specific seasons of the year, and field surveys should be performed during these seasons to identify important settlement areas. We specifically recommend that these often-missed first-stage habitats be considered in the seascape nursery concept.

Habitat connectivity: predictable diel, tidal and ontogenetic habitat shifts

Few species are confined to a single nursery habitat (Nagelkerken 2007). Seascape studies have shown that many animals utilize a mosaic of habitats on a daily basis (Boström *et al.* 2011). Mobile animals connect adjoining habitats through tidal, shelter-seeking or foraging movements (Hammerschlag *et al.* 2010; Igulu *et al.* 2013; Olds *et al.* 2013; Baker *et al.* 2013). These migrations are highly predictable in timing and routes followed (Krumme 2009), to such extent that some predators in nursery areas have adapted their behaviour

to coincide with these migrations (Helfman 1986). Animals pass through non-nursery habitats on a regular basis while moving between patches of core habitat in search of food or shelter (Hitt *et al.* 2011). These movements usually occur within a specified home range around the core area of their shelter sites (Farmer and Ault 2011), which are often located near to structurally complex habitats (Verweij and Nagelkerken 2007). Species often show homing behaviour to such shelter sites, which may persist over periods of weeks to months (Helfman *et al.* 1982). On longer time-scales, many species show ontogenetic shifts among habitats because of changing resource needs (e.g. food, shelter) as well as altered predation risk during different life stages (Dahlgren and Eggleston 2000; Kimirei *et al.* 2013b). Due to strong connectivity among habitat patches, assigning single nursery habitats disregards the role that earlier life-stage habitats or adjoining (feeding/shelter) habitats play in the population dynamics and ultimate stock replenishment of nursery species.

The seascape mosaic: hotspots of animal abundances and productivity

Spatially explicit use of patches within nursery habitats typically has not been quantified in relation to nursery function. In contrast, landscape-focused studies have demonstrated consistent and predictable animal density or productivity 'hotspots' in relation to spatial position within the seascape, for example based on: (i) distance to estuary mouth (Bell *et al.* 1988), (ii) distance to feeding areas (Pittman *et al.* 2007), (iii) proximity to high-volume tidal channels that supply larvae (Ford *et al.* 2010), (iv) density of creek edges within marshes (Kneib 2003), (v) presence and type of adjacent habitats (Nagelkerken *et al.* 2001), or (vi) specific salinity regimes representative of transitional areas between rivers and estuaries (Wasserman and Strydom 2011). Furthermore, habitat transition areas are specific zones within coastal seascapes that often have greater densities of organisms than areas further from edges (Dorenbosch *et al.* 2005). In many cases, the Beck *et al.* (2001) and Dahlgren *et al.* (2006) approaches may well identify the broad nursery habitat(s) used by a population, but miss critical mosaics of habitat patches in the seascape that underpin nursery function (Sheaves 2009).

Ecosystem corridors: highways connecting nurseries to adult populations

The last stage of nursery habitat occupancy during which organisms undertake their final migration to deeper or offshore waters to join the adult population is poorly known (Gillanders *et al.* 2003), but telemetry studies suggest that it can occur over short periods ranging from a few hours to days (Luo *et al.* 2009). Specific routes within estuaries or lagoons may act as preferred corridors that lower predation risk, span the shortest distance to reach deeper water or facilitate tidally enhanced movements due to specific local hydrology (Zollner and Lima 1999). Some studies have indicated the importance of continuous habitat edges (Hitt *et al.* 2011) or unvegetated strips within continuous seagrass beds as corridors (Bostrom *et al.* 2006), but extensive open shallow areas normally act as barriers for movement (Turgeon *et al.* 2010). In intertidal areas with extensive sand or mud flats, animals will often be funnelled to subtidal habitats through narrow tidal channels. From thereon, fish move to offshore waters by navigating through corridors such as deep channels, through narrow bay mouths or through open spaces among sandbanks, islets and other types of natural barriers situated at the ocean side of river deltas, estuaries and lagoons (e.g. Verweij *et al.* 2007; Luo *et al.* 2009). Incorporation of migration corridors and their temporal usage patterns is a critical consideration for the seascape nursery concept.

The seascape nursery: combining nursery-function and seascape-ecology concepts

Existing approaches to nursery habitat evaluation tend to give more weight to final juvenile stages prior to emigration to offshore adult stocks. Linkages among habitats that affect the critical growth and survival of earlier stages therefore tend to be underplayed. We suggest that the seascape nursery approach incorporates more fully those earlier stages. The importance of our approach is demonstrated in the following example for fishes with a complex life cycle. Consider a microtidal seascape (Fig. 1a) where fish settle largely in first-encountered, non-core habitats such as coral rubble areas along edges of tidal channels or at bay mouths, subsequently progress to seagrass beds, then switch to mangroves, and finally occupy hard-bottom

patch reefs or rocky areas, before moving to offshore reefs (example from Nagelkerken *et al.* 2000 and Grol *et al.* 2011). In this example, individuals are also found in other habitats, but those described above are where highest fish aggregation or production occurs. During seagrass and rubble occupancy, small juveniles feed and shelter in the same habitat to reduce predation, but at larger sizes, they use mangroves or patch reefs for shelter and show a diel or tidal migration to nearby seagrass beds to feed (Verweij *et al.* 2006). During these movements, they need to move from one feeding patch to another and pass through secondary habitats, such as algal beds and sand patches, which do not play an important role for feeding or as shelter but are part of their home range (see concentric circles in Fig. 1).

In the above example, the extensive seagrass beds provide the largest overall contribution to the adult populations (e.g. Verweij *et al.* 2008) and would be identified as the main nursery habitats

using the Dahlgren *et al.* (2006) approach. In contrast, expressed as a contribution per unit area, the importance of seagrass beds with large surface area would typically be lower compared with other habitats with smaller surface areas where crowding of animals occurs, such as mangrove stands and coral patches. Based on the Beck *et al.* (2001) approach, such habitats that contribute most per unit area could be designated as nursery habitats even though their overall production might not be large. This could in practice lead to a debate about whether mangroves vs. seagrass beds should be managed, what proportion of their total surface area should be conserved, and which areas within the estuary or lagoon should be managed, especially in cases of high usage or exploitation by multiple stakeholders. The seascape nursery would provide a more realistic approach to this problem by revealing that (Fig. 1) (i) transient settlement areas should be conserved, because without these, there is no recruitment to 'nursery' habitats; (ii)

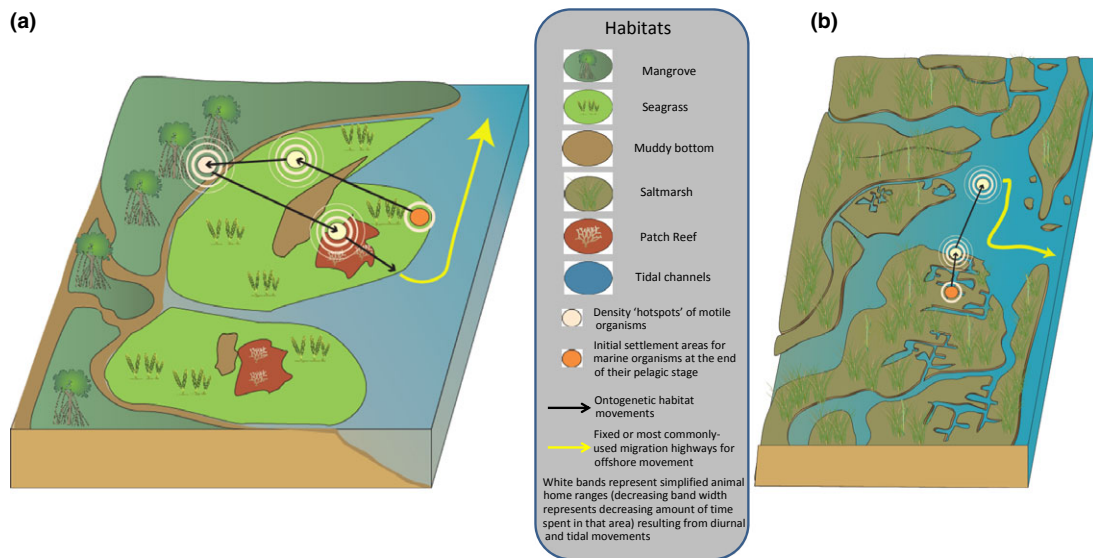


Figure 1 (a) Example as described in the text of a seascape nursery located in a clear water, microtidal lagoon supporting a variety of habitat types; the seascape nursery consists of several habitat mosaics connected through diel and ontogenetic movements. (b) Example of a seascape nursery for penaeid shrimps in a turbid, mesotidal salt marsh estuary. This specific case study refers to coastal salt marsh ecosystems of the northern Gulf of Mexico, which are considered critical in the support of highly productive shrimp fisheries (Turner 1977). Adults spawn offshore and post-larvae recruit to shallow habitats in the marsh complex of coastal bays and estuaries where conditions are favourable (salinity, temperature, food availability) (Rozas and Minello 2011). There is a staged ontogenetic progression of juveniles from the marsh complex to open bays, and subsequent migration to join adult stocks offshore (Lindner and Cook 1970). Although represented as circles for consistency of presentation, a narrow strip at the vegetation–open-water interface represents a density hotspot for juvenile shrimp within the marsh complex (Minello *et al.* 2008). Image credits: Kate Moore, Jane Thomas, Tracey Saxby and Diana Kleine (IAN Image Library – ian.umces.edu/imagelibrary) and Nina McLean (James Cook University).

within the seascape, there are principle areas (habitat mosaics), constrained by animal home ranges, that attract higher densities of mobile organisms and which are more productive than other areas, providing a management tool to prioritize areas of conservation; (iii) successive essential life-stage habitats should be conserved as impacts on one habitat affect productivity in habitats occupied during later life stages; (iv) without conserving migration routes that connect different animal hotspots during ontogeny or that facilitate movement from nurseries to offshore populations, nurseries could experience a switch from acting as sources to becoming juvenile sinks. A similar example from a mesotidal salt marsh system is provided in Fig. 1B.

Not all species show a complex life cycle such as described above. Nevertheless, it is a common observation for a multitude of species that tidal channels are favoured for movement through shallow areas, that animal abundances are highly correlated with spatial position within coastal habitats (e.g. driven by salinity or turbidity gradients) and that animals regularly perform diel or tidal movements (Whaley *et al.* 2007; Krumme 2009; Turgeon *et al.* 2010). So even for species with a relatively simple life cycle, in terms of habitat use, previous approaches fail to incorporate several important dynamic processes other than ontogenetic habitat shifts.

Practical steps to seascape nursery analysis

While there is no single best approach to identify the precise mosaic of habitats most essential during the juvenile stages of animals in coastal marine environments, it is crucial to recognize the importance of a mosaic of contributing habitats and their linkages. Here, we outline the practical steps that can help improve on earlier approaches for identification and evaluation of nursery habitat and ultimately lead to more successful protection and management of nursery function. The order and relative importance of these steps will vary depending on specific situations.

Step 1

Following Beck *et al.* (2001), identify the relative contribution to adult populations of all juvenile habitats at whatever scale they can be identified, for example, using approaches such as otolith

microchemistry (Gillanders and Kingsford 1996; Verweij *et al.* 2008). This will typically be at a coarser scale than relevant to management objectives (e.g. whole estuary or whole habitat unit) and fail to identify linkages across the seascape. We therefore recommend subsequent work to identify the smaller-scale patches within each broad-scale nursery that contribute most to the overall population replenishment by that nursery habitat. This will likely, but not necessarily overlap with density hotspots of juvenile animals during their inactive as well as active period (e.g. Ford *et al.* 2010), which can be identified through field surveys. Identification of specific patches that contribute most to the overall production of a nursery habitat is more challenging, but techniques such as stable isotope analysis of muscle tissue, internal and external artificial tags, or genetic and chemical markers can provide the necessary finer-scale information (Gillanders 2009; Kimirei *et al.* 2013a), as well as provide an answer to how this contribution may vary over time (see e.g. Kraus and Secor 2005).

Step 2

Known (from the literature) or field-acquired (through tagging studies) home-range sizes may then be projected onto the identified highest-productivity density hotspots to establish the effective area that is used as a juvenile habitat (the habitat mosaic). The home range includes the seascape that is most used on a daily basis for activities such as sheltering and foraging. Home-range sizes around hotspots of animal abundances could be considered at decreasing levels of importance (see Fig. 1). Using radii of these dimensions should prove to be a more effective way to manage nursery mosaics than a static approach of single complete habitats because it uses broader information on critical habitat use. While tagging juvenile animals is difficult and movement ranges can differ considerably among species and within habitats, home-range size is often a function of body size (Kramer and Chapman 1999) and juveniles of most demersal species show high site fidelity and restrict their movements to distances of no more than a few 100s m from their preferred shelter sites (Tupper 2007; Nagelkerken *et al.* 2008). Home ranges are larger in cases where animals occupy macrotidal habitats, but also in this case, fidelity has been shown to high-tide and low-tide

habitat components (Dorenbosch *et al.* 2004; Hering *et al.* 2010).

Step 3

Patterns of ontogenetic habitat shifts should be identified for animals that occupy the above high-productivity hotspots, so that other habitat patches that are previously or subsequently occupied are included in the designation of effective nursery mosaic (Fig. 1). This is based on the principle that patches that contribute most to adult populations can only sustain this productivity as a result of habitat linkages through ontogeny. Approaches such as following the progression of cohorts (abundances and sizes of organisms) in multiple juvenile habitats can identify which habitats are most likely to play a key role in provisioning recruits to next life-stage habitats (e.g. Fodrie *et al.* 2009). A critical consideration in this is to identify primary settlement areas where early life stages occur, typically at sizes at which they have not been included in field surveys.

Step 4

Primary migration routes should be identified (e.g. using telemetry or conventional tagging) that connect animal production hotspots across different spatiotemporal scales. This includes corridors that facilitate animal movement from one habitat mosaic to another through ontogeny, as well as from the seascape nursery to offshore waters (Fig. 1). Migration highways are likely to overlap among species based on the same advantages that they provide for a suite of species, such as structure-rich corridors that facilitate movement under lowered predation risk (Gilliam and Fraser 2001). In deep-water estuaries and lagoons, such migration corridors might be less evident or relevant than in shallower ecosystems dominated by extensive mud or sand flats. However, due to the geomorphology of many inshore water bodies around the world, animals still need to pass through bay mouths, openings between barrier islets, or through deeper tidal channels to reach offshore waters. As such, these areas should be given high conservation importance as they maintain connectivity among inshore and offshore ecosystems.

We have attempted to present an improved framework to identify nurseries for management

purposes that we believe will provide an acceptable level of accuracy for a wide range of species in a variety of coastal marine ecosystems. Our approach does not provide a single, best solution for multispecies management, as different groups of species may occupy different combinations of habitats or different areas of estuaries and lagoons. As is the case for previous approaches of nursery identification, trade-offs need to be made in terms of which species and which areas receive most consideration in terms of conservation or management. While for some systems with few, highly abundant fishery species and just one or two habitat types, a coarse approach such as that of Dahlgren *et al.* (2006) and Beck *et al.* (2001) may provide a reasonable amount of information for management purposes, there are many other systems and a multitude of (commercial and keystone) species where such an approach is likely to fail. The seascape nursery approach adds more realism to the identification of core juvenile areas within these systems by incorporating spatiotemporal drivers of animal habitat use. The intention is to achieve a practical advance for the conservation and management of inshore coastal areas that are highly productive for coastal fisheries but also prone to high levels of competing demands and degradation through human activities. We also recommend consideration of more challenging, dynamic management approaches such as mobile protected areas that follow movements of key species across their landscape (Bull *et al.* 2013).

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